# Spatial and Temporal Disparity of Fish Assemblage Relationship with Hydrological Factors in Two Rivers Tangon and Kulik, Thakurgaon, Bangladesh 

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#### Abstract

Stream fish assemblage and their association with major ecological factors were spatiotemporally conducted from January to December 2015 at the Tangon and Kulik Rivers of Thakurgaon district in Bangladesh. A total of 6,561 specimens belonging to 53 fish species dominated (>4.86\%) by Aspidoparia jaya, Pethia ticto, Puntius sophore, Channa punctatus and Canthophrys gongota both in rivers and seasons. Significant differences ( $\mathrm{P}<0.05$ ) were spatiotemporally observed in the values of diversity indices between two rivers. Spatial and temporal patterns influenced species copiousness and richness that higher in winter season due to little water echelon. Fish communities were differed between rivers $(R=0.46, \mathrm{P}<0.01)$ and seasons ( $R=0.21, \mathrm{P}<0.01$ ) stressing as 0.24 and dividing into two clusters at a value of $87.65 \%$ union. Canonical Correspondence Analysis (CCA) specified that water depth, water temperature and transparency were key environmental factors affecting fish assembly, abundance and distribution. Therefore, advanced plans and management strategies should be taken to save threatened fishes at low water depth during dry and winter seasons than others.


Keywords: Spatiotemporal discrepancy, fish diversity indices, ecological factors, rivers.

## Introduction

Aquatic ecologists has great interest to known the assemblages of fishes including their changing patterns in abundance and frequency familiar as imperative tools of fisheries management and conservation. Riverine or stream fishes of the developing world have faced to different ecological stresses affecting fish assemblage structure of that habitat (Terra et al., 2016). However, association between fishes and their environments play central roles for managing and saving of riverine species where any modification of it can lead to transform in their population arrangement (Kadye and Moyo, 2007). The concentrations of environmental factors mainly water quality parameters have been accounted to persuade assemblage and distribution of fishes both in freshwater and marine networks (Tunesi et al., 2006; Hossain et al., 2012; Corpuz et al., 2015). Biodiversity indicators i.e. dominance, evenness, Margalef and Shannon-Weiner diversity indices has used as pointer to discern the assortment status of aquatic residents (Magurran, 1988; Vyas et al., 2012).

In Bangladesh, aquatic habitats gradually condense fish abundance and species (Chaki et al., 2014), have endured and abused through extreme
human interference (Hossain et al., 2012). The Tangon and Kulik Rivers are said to be lifeline in Thakurgaon district northern part of Bangladesh but a feeble flow is now flowing in the middle of the rivers during dry and winter month. Due to siltation and petite water level, a number of freshwater fish species specially threatened one are gradually declined their abundance each year (Rahman et al., 2003). So, a management plan and strategy should be taken to conserve fishes in this river. The central views of this research are firstly aimed 1) to explain the spatiotemporal variation of fish assemblage and 2) to detect the outline of relationship between fish grouping and major hydrological factors at the Tangon and Kulik Rivers of Bangladesh.

## Materials and Methods

## Study area

A study was planned at the Tangon and Kulik Rivers (Figure 1) in Thakurgaon district of Bangladesh from January to December 2015. Three months were considered as one season based on their similarities i.e. February, March and April as premonsoon (prm); May, June and July as monsoon


Figure 1. GIS location of sampling sites in the Tangon (St.1, St. 2 and St.3) and Kulik (St.4, St. 5 and St.6) Rivers during four seasons (pre-monsoon, monsoon, post-monsoon and winter season).
(ms); August, September and October as postmonsoon (pom); and November, December and January as winter (wm) season. Randomly six stations were selected having about 5 km distances from each other where station $1\left(25.88751^{\circ} \mathrm{N} 88.40760^{\circ} \mathrm{E}\right.$, altitude 41.50 m$), 2\left(25.85501^{\circ} \mathrm{N} 88.40127^{\circ} \mathrm{E}, 43.60\right.$ $\mathrm{m})$ and $3\left(25.82056^{\circ} \mathrm{N} 88.38496^{\circ} \mathrm{E}, 40.40 \mathrm{~m}\right)$ in the Tangon River while stations 4 (25.91520 ${ }^{\circ} \mathrm{N}$ $\left.88.27375^{\circ} \mathrm{E}, 46.20 \mathrm{~m}\right), 5\left(25.88960^{\circ} \mathrm{N} 88.26985^{\circ} \mathrm{E}\right.$, $43.70 \mathrm{~m})$ and $6\left(25.86703^{\circ} \mathrm{N} 88.25061^{\circ} \mathrm{E}, 42.40 \mathrm{~m}\right)$ in the Kulik River.

## Sample Collection

During study period, data were collected at monthly intervals for water quality parameters and fish species. Water temperature ( ${ }^{\circ} \mathrm{C}$ ), water depth (m), dissolved oxygen ( $\mathrm{mg} / \mathrm{l}$ ), water pH and transparency (cm) were measured by a standard method (APHA, 2012) in-situ using a digital thermometer (Digithermo), DO meter (Model: DO5509, Lutron), depth meter (traditional), pH meter (Model: HI-8014, HANNA instruments) and Secchi disk, respectively. Fish specimens were directly collected by fishermen where fishes were caught with seine net $\left(15 \times 3.5 \mathrm{~m}^{2}\right.$, mesh size 4 mm$)$ and cast net ( $3.5 \times 6.5 \mathrm{~m}^{2}, 8 \mathrm{~mm}$ ). These fishing gears were operated at same sampling
spot within 0.5 km area to ensure maximum harvesting of fishes. At each station (8.00 AM), five throws were made for cast net and three hauls for seine net to catch fishes. Sampling that less than two species per throws and hauls were kept out for counting. Then, collected fishes were sorted based on their key morphological characters and counted their individual abundance. Species that seemed difficult to identify on spot were preserved in 7 to $10 \%$ buffered formalin solution and conveyed to Laboratory of the Department of Fisheries Biology and Genetics under Hajee Mohammad Danesh Science and Technology University (HSTU) (Bangladesh) to facilitate identification and further study. The ichthyo-faunas were systematically identified and classified based on their morphology followed by Rahman (2005).

## Statistical Analyses

In case of Kulik River, three months explicitly March, April and May known as dry season were excluded from analysis because of very low water depth ( $<0.30 \mathrm{~m}$ ) or somewhere dry out. For hydrological parameters (temperature, dissolved oxygen, pH and transparency) grouped by space and time, one-way analysis of variance (ANOVA) followed by Tukey's test were used to definite
dissimilarities among between rivers and seasons. Canonical Correspondence Analysis (CCA) was designed to explore the correspondence between physical factors and species compositions, and to know the fish communities (Toham and Teugels, 1998). To assess the relative importance of each hydrological variable, we used CCA on each river and season derived from abundance and hydrological matrices. CCA was applied to overall fish data matrix and environmental data matrix obtaining a direct environmental interpretation of extracted ordination axes. To know the sample adequacy both in space and season, species accumulation curves were brought into play by AccuCurve version 1.0 (Drozd and Novotny, 2010). Four major biodiversity indices namely dominance, evenness, Margalef richness and Shannon-Weiner may used to know the discrepancy of aquatic communities or populations. However, to know the status of fish community structure and assembly, data were monthly collected and traced where diversity indices were calculated the as BuzasGibson's evenness $E=e^{H} / S$ (Pielou, 1966), Dominance index $D=\sum_{i}\left(\frac{n_{i}}{n}\right)^{2}$ (Harper, 1999), Margalef's richness index $d=(S-1) / \ln (n)$ (Margalef, 1968) and Shannon-Weiner diversity index $H=-\sum_{i} \frac{n_{i}}{n} \ln \frac{n_{i}}{n}$ (Shannon \& Weiner, 1949). Where, $n_{i}$ is the number of individuals of taxon $i ; n$ is total number of individuals; $\ln$ is natural logarithm; $S$ is number of taxa; $e$ is natural logarithm equal to 2.718. The irregular and scarce fish species (> 2 individuals) in both rivers were not considered for multivariate analysis. Using two-dimensional nonmetric multidimensional scaling (nMDS), fish abundance data were transforming as $\ln (x+1)$ in order to moderate influence of extreme values and fish assemblage were reviewed as spatial and temporal scales. Based on Bray-Curtis similarity method, major contributory fishes that responsible for parallel in grouping were dogged with SIMPER analysis (similarity percentage). Besides, one-way analysis of similarities (ANOSIM) was tested to assess the significant variations in spatiotemporal scales. A nonparametric test known as PERMANOVA was also used to compare the fish abundance data between rivers and seasons, respectively. An affiliation among
fish assemblages from each station and month were graphically compared through cluster analysis by Unweighted Pair Group Method with Arithmetic mean (UPGMA) by Clarke and Warwick (1994). All statistical analysis was done using PAST (Paleontological Statistics, version 2.17 and 3.10) software (Hammer et al., 2001).

## Results

## Water quality parameters

Average values of major hydrological factors from two rivers and four seasons are presented in Table 1. No considerable differences $(P>0.05)$ were observed in water temperature $(F=0.02), \mathrm{pH}(F=$ 2.29) and transparency ( $F=0.09$ ) between Tangon and Kulik Rivers while significant variations ( $P$ < 0.05 ) were found in dissolved oxygen ( $F=6.23$ ) and water depth $(F=14.73)$, respectively. In contrast, significant differences were found in water temperature ( $F=43.73, P<0.01$ ), dissolved oxygen $(F=6.66, P<0.01), \mathrm{pH}(F=7.68, P<0.01)$, transparency $(F=13.57, P<0.01)$ and water depth $(F$ $=4.45, P<0.05)$ among seasons.

## Fishes and Water Quality Parameters

In case of water quality parameters, eigen values of canonical correspondence analysis (CCA) of the first four axes were found to be $0.21\left(\mathrm{CCA}_{1}\right), 0.15$ $\left(\mathrm{CCA}_{2}\right), 0.06\left(\mathrm{CCA}_{3}\right)$ and $0.04\left(\mathrm{CCA}_{4}\right)$ both for spatial and temporal scale where first $\left(\mathrm{CCA}_{1}\right)$ and second $\left(\mathrm{CCA}_{2}\right)$ axes were pooled and modeled as $45.73 \%$ and $32.95 \%$ of species data, respectively (Figure 2). Vector length of any specific parameters is a sign of magnitude of that variable in CCA analysis. The highest vector length of water depth at fourth axis showed significant correlation with the occurrence of Cirrhinus reba, Gagata youssoufi and Xenentodon cancila in post-monsoon and winter season at the Tangon and Kulik Rivers, respectively. Besides, vector length of water temperature was also correlated with the profusion of Channa punctatus and Clarias batrachus at both rivers in post-monsoon but at the Kulik River in winter months. The vector of pH was associated with the abundance of Anabas tenstudineus, Aspidoparia jaya and Pethia ticto in

Table 1. Average value $( \pm \mathrm{SD})$ of major water quality parameters with their variation $(P<0.05)$ in both rivers and seasons

| Hydrological factors | Rivers |  |  | Seasons |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Tangon | Kulik |  | prm | ms | pom | wm |
| Water temperature $\left({ }^{\circ} \mathrm{C}\right)$ | $27.84 \pm 4.86$ | $27.64 \pm 5.02$ |  | $24.21 \pm 2.70^{\mathrm{a}}$ | $31.94 \pm 2.08^{\mathrm{b}}$ | $31.27 \pm 2.05^{\mathrm{b}}$ | $23.11 \pm 3.82^{\mathrm{a}}$ |
| Water depth $(\mathrm{m})$ | $2.00 \pm 0.64^{\mathrm{a}}$ | $1.47 \pm 0.36^{\mathrm{b}}$ |  | $1.52 \pm 0.39^{\mathrm{a}}$ | $2.15 \pm 0.94^{\mathrm{b}}$ | $1.83 \pm 0.45$ | $1.43 \pm 0.60$ |
| Dissolved oxygen $(\mathrm{mg} / \mathrm{l})$ | $6.35 \pm 0.98^{\mathrm{a}}$ | $5.76 \pm 0.83^{\mathrm{b}}$ |  | $5.43 \pm 0.9^{\mathrm{a}}$ | $5.73 \pm 1.04^{\mathrm{a}}$ | $6.71 \pm 0.82^{\mathrm{b}}$ | $6.24 \pm 0.55$ |
| Water pH | $7.27 \pm 0.41$ | $7.09 \pm 0.55$ |  | $7.00 \pm 0.47^{\mathrm{a}}$ | $6.95 \pm 0.33^{\mathrm{a}}$ | $7.58 \pm 0.18^{\mathrm{b}}$ | $7.14 \pm 0.58^{\mathrm{a}}$ |
| Transparency $(\mathrm{cm})$ | $29.58 \pm 3.10$ | $29.37 \pm 2.53$ |  | $27.63 \pm 2.97^{\mathrm{a}}$ | $31.11 \pm 1.62^{\mathrm{b}}$ | $31.30 \pm 1.33^{\mathrm{b}}$ | $27.58 \pm 2.81^{\mathrm{a}}$ |

SD, standard deviation; prm, pre-monsoon; ms, monsoon; pom, post-monsoon; wm, winter; values of the parameter in each row with different superscripts ( $\mathrm{a}, \mathrm{b}$ and c ) differs significantly ( $\mathrm{P}<0.05$ )


Figure 2. Canonical correspondence analysis (CCA) of fish abundance (S, species) and hydrological parameters (WT, water temperature; DO, dissolved oxygen; pH , water pH ; TR, transparency) of the Tangon ( $\bullet$ ) and Kulik (+) Rivers during different seasons (prm, premonsoon; ms, monsoon; pom, post monsoon; wm, winter).
winter season at the Kulik River while in both premonsoon and monsoon at the Tangon River. The vector length of dissolved oxygen was only connected with the occurrence of G. youssoufi in winter season at the Kulik River where transparency with $C$. punctatus and Macrognathus pancalus in all seasons at the Kulik River.

## Diversity Indices of Fishes

A total of 6,561 fish specimens, belonging to 38 genera and 53 fish species including 14 threatened species (Table 2), were collected both from the Tangon ( $82.73 \%$ ) and Kulik ( $17.27 \%$ ) Rivers. A species horizontal accumulation curve was arrived at an asymptote where fishes were accrued with a parallel and growing pattern leveling the effects of spatial and temporal heterogeneity both in Tangon and Kulik Rivers (Figure 3). Based on space and time, common values (mean $\pm$ SD) of diversity indices from rivers and seasons are given in Table 3. Significant differences ( $P<0.01$ ) were spatially observed in the values of dominance ( $F=7.57$ ), evenness ( $F=10.72$ ) and Shannon ( $F=8.55$,) among two rivers except for Margalef ( $F=1.84, P>0.05$ ). On the contrary, significant differences $(P<0.01)$ were also determined in the values of dominance ( $F=14.30$ ), evenness ( $F=3.80$ ), Margalef $(F=27.89)$ and Shannon ( $F=22.38$ ) among season from these rivers.

## Assembly and Structure of Fishes

Based on SIMPER analysis (all pooling), about $83.73 \%$ and $79.19 \%$ average dissimilarity were found in rivers and seasons, respectively (Table 4). The
highest contributing species were A. jaya ( $10.13 \%$ and $9.76 \%$ ), P. ticto ( $7.77 \%$ and $7.79 \%$ ), Puntius sophore ( $5.76 \%$ and $5.25 \%$ ), C. gongota ( $5.13 \%$ and $4.86 \%$ ) and C. punctatus ( $5.12 \%$ and $4.94 \%$ ) while lowest was Nandus nandus ( $0.24 \%$ and $0.27 \%$ ) both in spatiotemporal scales, respectively. A twodimensional nMDS based on Bray-Curtis's similarity index suggests that fish assemblages at the Tangon River was speckled from Kulik River while monsoon was separated from another seasons pressuring as 0.24 (Figure 4 and 5). Analysis of similarity (ANOSIM) showed considerable variation in fish grouping at the rivers and seasons. The Tangon River showed significant difference ( $R=0.46, P<0.01$ ) in fish assemblage with the Kulik River. Moreover, fish assemblage were significantly different ( $R=0.21, P<$ 0.01 ) among season where highest contribution was found in winter but lowest in monsoon season. Based on Euclidean method, a non-parametric PERMANOVA test also exhibited significant variations ( $P<0.01$ ) in individual assemblage data between rivers $(F=8.80)$ and seasons ( $F=3.48$ ), respectively. Two clusters were spatiotemporally perceived as $12.35 \%$ partition using Bray-Curtis's similarity index where (Figure 6) the Kulik River was separated from the Tangon River in monsoon from other seasons. Besides, two sub-clusters were also viewed where $1^{\text {st }}$ sub-cluster in monsoon months at the Tangon River which was also isolated from another season due to very close fish resemblance.

## Discussion

Aquatic biodiversity of a stream is strongly influenced by water chemistry and habitat quality

Table 2. Spatiotemporal abundance and allocation of fishes from rivers and seasons including their present status

| Genus name | Scientific name | Fish code | IUCN <br> status | Rivers |  | Contribution (\%) | Seasons |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | TR | KR |  | prm | ms | pom | wm |
| Acanthocobitis | Acanthocobitis botia | S1 | LC | 134 | 18 | 2.32 | 12 | 58 | 20 | 62 |
| Amblypharyngodon | Amblypharyngodon mola | S2 | LC | 30 | 42 | 1.10 | 8 | 2 | 29 | 33 |
| Anabas | Anabas testudineus | S3 | LC | 60 | 0 | 0.91 | 16 | 8 | 16 | 20 |
| Aspidoparia | Aspidoparia jaya | S4 | LC | 690 | 103 | 12.09 | 51 | 120 | 261 | 361 |
|  | Aspidoparia morar* | S5 | VU | 90 | 0 | 1.37 | 0 | 0 | 90 | 0 |
| Barilius | Barilius barna* | S6 | EN | 234 | 59 | 4.46 | 12 | 42 | 180 | 59 |
|  | Barilius bendelisis* | S7 | EN | 42 | 0 | 0.64 | 0 | 0 | 0 | 42 |
| Botia | Botia lohachata* | S8 | EN | 84 | 15 | 1.51 | 24 | 30 | 0 | 45 |
| Canthophrys | Canthophrys gongota | S9 | NT | 306 | 36 | 5.21 | 90 | 24 | 12 | 216 |
| Chagunius | Chagunius chagunio* | S10 | VU | 60 | 18 | 1.19 | 0 | 0 | 78 | 0 |
| Chanda | Chanda nama | S11 | LC | 120 | 0 | 1.83 | 120 | 0 | 0 | 0 |
|  | Channa orientalis | S12 | LC | 60 | 0 | 0.91 | 0 | 0 | 0 | 60 |
| Channa | Channa punctatus | S13 | LC | 330 | 24 | 5.40 | 24 | 0 | 219 | 111 |
|  | Channa striatus | S14 | LC | 48 | 18 | 1.01 | 48 | 0 | 9 | 9 |
| Cirrhinus | Cirrhinus cirrhosus | S15 | NT | 96 | 0 | 1.46 | 30 | 0 | 0 | 66 |
|  | Cirrhinus reba | S16 | NT | 78 | 6 | 1.28 | 12 | 12 | 33 | 27 |
| Clarias | Clarias batrachus | S17 | LC | 40 | 12 | 0.79 | 16 | 8 | 11 | 17 |
| Crossocheilus | Crossocheilus latius* | S18 | EN | 30 | 0 | 0.46 | 12 | 0 | 0 | 18 |
| Devario | Devario devario | S19 | LC | 18 | 54 | 1.097 | 9 | 0 | 45 | 18 |
| Erethistes | Erethistes pussilus | S20 | LC | 22 | 0 | 0.34 | 0 | 0 | 0 | 22 |
| Esomus | Esomus danricus | S21 | LC | 78 | 0 | 1.19 | 0 | 0 | 0 | 78 |
| Eutropiichthys | Eutropiichthys vacha | S22 | LC | 20 | 6 | 0.40 | 0 | 0 | 10 | 16 |
| Gagata | Gagata cenia | S23 | LC | 110 | 23 | 2.03 | 27 | 14 | 35 | 57 |
|  | Gagata youssoufi | S24 | NT | 82 | 21 | 1.57 | 14 | 16 | 40 | 33 |
| Glossogobius | Glossogobius giuris | S25 | LC | 180 | 21 | 3.06 | 48 | 42 | 30 | 81 |
| Heteropneustes | Heteropneustes fossilis | S26 | LC | 42 | 0 | 0.64 | 0 | 0 | 42 | 0 |
|  | Labeo angra | S27 | LC | 62 | 10 | 1.10 | 0 | 0 | 7 | 65 |
|  | Labeo bata | S28 | LC | 72 | 6 | 1.19 | 0 | 51 | 18 | 9 |
| Labeo | Labeo boga* | S29 | CR | 20 | 4 | 0.36 | 0 | 0 | 0 | 24 |
|  | Labeo calbasu | S30 | LC | 18 | 0 | 0.27 | 10 | 0 | 2 | 6 |
|  | Labeo gonius | S31 | NT | 26 | 2 | 0.43 | 4 | 16 | 0 | 8 |
|  | Labeo rohita | S32 | LC | 138 | 6 | 2.19 | 30 | 48 | 9 | 57 |
| $\begin{aligned} & \text { Lepidocephalichthy } \\ & \text { s } \end{aligned}$ | Lepidocephalichthys guntea | S33 | LC | 120 | 30 | 2.29 | 42 | 66 | 39 | 3 |
| Macrognathus | Macrognathus aculeatus | S34 | NT | 48 | 0 | 0.73 | 48 | 0 | 0 | 0 |
|  | Macrognathus pancalus | S35 | LC | 240 | 71 | 4.74 | 9 | 78 | 113 | 111 |
| Mastacembalus | Mastacembalus armatus* | S36 | EN | 60 | 0 | 0.91 | 0 | 0 | 48 | 12 |
| Mystus | Mystus bleekeri | S37 | LC | 20 | 4 | 0.37 | 0 | 0 | 2 | 22 |
|  | Mystus tengara | S38 | LC | 138 | 11 | 2.27 | 12 | 42 | 20 | 75 |
| Nandus | Nandus nandus | S39 | NT | 14 | 3 | 0.26 | 0 | 0 | 0 | 17 |
| Ompok | Ompok pabda* | S40 | EN | 20 | 3 | 0.35 | 0 | 12 | 8 | 3 |
| Pethia | Pethia ticto* | S41 | VU | 558 | 138 | 10.61 | 141 | 72 | 189 | 294 |
| Genus name | Scientific name | Fish | IUCN | Rivers |  | Contribution | Seasons |  |  |  |
|  |  | code | status | TR | KR | (\%) | prm | ms | pom | wm |
| Pseudambassis | Pseudambassis lala | S42 | LC | 18 | 9 | 0.41 | 0 | 0 | 0 | 27 |
| Psilorhynchus | Psilorhynchus balitora | S43 | LC | 18 | 3 | 0.32 | 0 | 12 | 0 | 9 |
| Puntius | Puntius sophore | S44 | LC | 264 | 48 | 4.76 | 72 | 132 | 87 | 21 |
| Raiamas | Raiamas bola* | S45 | EN | 24 | 0 | 0.37 | 0 | 0 | 6 | 18 |
| Rita | Rita rita* | S46 | EN | 8 | 9 | 0.26 | 8 | 0 | 0 | 9 |
| Salmophasia | Salmophasia bacaila | S47 | LC | 108 | 200 | 4.69 | 66 | 49 | 87 | 106 |
| Salmostoma | Salmostoma phulo | S48 | NT | 48 | 0 | 0.73 | 0 | 0 | 0 | 48 |
| Sperata | Sperata aor* | S49 | VU | 150 | 52 | 3.08 | 3 | 38 | 156 | 5 |
| Tetraodon | Tetraodon cutcutia | S50 | LC | 42 | 0 | 0.64 | 0 | 0 | 0 | 42 |
| Trichogaster | Trichogaster fasciata | S51 | LC | 64 | 3 | 1.02 | 0 | 0 | 40 | 27 |
| Wallago | Wallago attu* | S52 | VU | 11 | 7 | 0.27 | 0 | 7 | 6 | 5 |
| Xenentodon | Xenentodon cancila | S53 | LC | 42 | 31 | 1.11 | 13 | 12 | 28 | 20 |
| Total Contribution (\%) |  |  |  | 5435 | 1126 | 100 | $1031$ | $1011$ | 2025 | 2494 |
|  |  |  |  | 82.84 | 17.16 | 100 | 15.71 | 15.41 | 30.86 | 38.01 |

$\overline{T R}$, Tangon River; KR, Kulik River; prm, pre-monsoon; ms, monsoon; pom, post monsoon; wm, winter; *Threatened fishes; IUCN: International Union for Conservation of Nature; $C R$ : Critically endangered; $E N$ : Endangered; $V U$ : vulnerable; $N T$ : Near threatened; $L C$ : Least concern


Figure 3. A species accumulation curve at the Tangon and Kulik Rivers (R) during four seasons (S).

Table 3. Descriptive statistics with one way ANOVA of diversity indices at the Tangon and Kulik Rivers during different seasons

| Rivers/seasons | Statistics | Diversity indices |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Dominance | Evenness | Margalef | Shannon |
| Rivers Tangon | Min | 0.07 | 0.69 | 0.83 | 1.52 |
|  | Max | 0.23 | 0.94 | 3.61 | 2.75 |
|  | Mean $\pm$ SD | $0.13 \pm 0.04^{\text {a }}$ | $0.81 \pm 0.07$ | $2.23 \pm 0.76$ | $2.22 \pm 0.35$ |
|  | Min | 0.09 | 0.69 | 0.42 | 0.47 |
| Seasons | Max | 0.70 | 0.97 | 3.09 | 2.48 |
|  | Mean $\pm$ SD | $0.23 \pm 0.19^{\text {b }}$ | $0.87 \pm 0.07$ | $1.96 \pm 0.84$ | $1.84 \pm 0.65$ |
|  | Min | 0.08 | 0.81 | 1.02 | 1.52 |
|  | Max | 0.23 | 0.97 | 3.52 | 2.73 |
|  | Mean $\pm$ SD | $0.13 \pm 0.04^{\text {a }}$ | $0.88 \pm 0.06^{\text {a }}$ | $2.34 \pm 0.77^{\text {a }}$ | $2.21 \pm 0.37^{\text {a }}$ |
|  | Min | 0.15 | 0.79 | 0.42 | 0.47 |
|  | Max | 0.70 | 0.96 | 1.92 | 2.06 |
|  | Mean $\pm$ SD | $0.33 \pm 0.2^{\text {b }}$ | $0.86 \pm 0.05$ | $1.05 \pm 0.49^{\text {b }}$ | $1.38 \pm 0.59^{\text {b }}$ |
|  | Min | 0.09 | 0.69 | 1.82 | 1.92 |
|  | Max | 0.17 | 0.95 | 2.98 | 2.53 |
|  | Mean $\pm$ SD | $0.13 \pm 0.03^{\text {a }}$ | $0.81 \pm 0.09^{\text {b }}$ | $2.37 \pm 0.39^{\text {a }}$ | $2.22 \pm 0.18^{\text {a }}$ |
|  | Min | 0.07 | 0.69 | 1.89 | 1.82 |
|  | Max | 0.21 | 0.93 | 3.61 | 2.75 |
|  | Mean $\pm$ SD | $0.12 \pm 0.04$ | $0.81 \pm 0.07^{\text {b }}$ | $2.61 \pm 0.48^{\text {a }}$ | $2.35 \pm 0.26^{\text {a }}$ |

prm, pre-monsoon; ms, monsoon; pom, post-monsoon; wm, winter; min, minimum; max, maximum; SD, standard deviation; values of the parameter in each column with different superscripts ( $\mathrm{a}, \mathrm{b}$ and c ) differs significantly $(P<0.05)$

Table 4. Discriminating contribution of fishes (>2.0\%) through SIMPER analysis in rivers and seasons

| Rivers $(83.73 \%)$ | Average dissimilarity | Seasons $(79.19 \%)$ |
| :--- | :---: | :---: |
| Contribution $(\%)$ | Major contributory fishes | Contribution $(\%)$ |
| 10.13 | Aspidoparia joya | 9.76 |
| 7.77 | Pethia ticto | 7.79 |
| 5.76 | Puntius sophore | 5.25 |
| 5.13 | Canthophrys gongota | 4.86 |
| 5.12 | Channa punctatus | 4.94 |
| 4.40 | Macrognathus pancalus | 4.56 |
| 4.38 | Salmophasia bacaila | 4.33 |
| 3.77 | Barilius barna | 4.01 |
| 3.33 | Sperata aor | 3.79 |
| 3.20 | Lepidocephalichthys guntea | 3.23 |
| 3.18 | Glossogobius giuris | 3.00 |
| 2.68 | Labeo rohita | 2.37 |
| 2.58 | Mystus tengara | 2.38 |
| 2.54 | Chanda nama | 2.56 |
| 2.40 | Acanthocobitis botia | 2.35 |



Figure 5. Two dimensional nMDS scaling of comparative fish assemblage data among seasons (prm, pre-monsoon; ms, monsoon; pom, post monsoon; wm, winter) in the Tangon and Kulik Rivers stressing as 0.24 .


Figure 6. Classical UPGMA clustering based on Bray-Curtis's similarity index of fish assemblage both in space (TR, Tangon River; KR, Kulik River) and time (prm, pre-monsoon; ms, monsoon; pom, post monsoon; wm, winter).
(Bio et al., 2011). Variations of water temperature among seasons showed more impacts on species distribution in Kulik River may be due to water depth than Tangon River. From these rivers, maximum level of fish abundance was detected at low water temperature and small flowing discharge in winter but minimum fish diversity viewed with comparatively high temperature and water discharge in summer differed by Yan et al. (2010) due to geographical variation. Alteration in water quality parameters such as water temperature, dissolved oxygen, pH , transparency and depth influence the uniqueness of aquatic environment and fish breeding (Kathiresan and Bingham, 2001; Rashleigh, 2004), profusion and
allotment (Maes et al., 2004), migration and distribution (Vega-Cendejas et al., 2013) and survival of fishes (Whitfield, 1999) ultimately altering fish assemblage and structure. However, significant differences ( $P<0.05$ ) were observed in hydrological parameters among habitat and season similar to Grimaldo et al. (2012). Besides, values of water quality parameters from these rivers were within the limits reported by Rakiba and Ferdoushi (2013) due to close ecological area. Fish assemblage and structure are mainly affected by seasonal changes along with ecological factors in estuaries (Loneragan and Potter, 1990; Young and Potter, 2003).

The number of fishes and individuals, native and
susceptible species, were found more from Tangon River than Kulik River over the successive seasons might be due to more periphyton community, charitable shelters, food staffs and breeding places. Quantity of fish species and specimens may vary due to scattering from the Tangon River in early April for breeding purposes. Maximum number of individuals was observed at Tangon River may be as a result of most favorable environmental condition mainly for water depth than Kulik River. From both rivers, highest number of species and specimens were caught in winter season may be due to reduced volume of water. Minimum fish species and specimen were recorded in monsoon would be due to low effectiveness of fishing gears used at higher depth as a result of heavy rainfall and floods during this period. The highest abundance of fishes recorded in November but lowest in June and August from Padma River (Chaki et al., 2014) and was deviated with Jahan et al. (2014) loads cresting in October and lowing in February may be because of geographical and environmental variations. An asymptote species accumulation curve attained for sampling sufficiency supported by Malcolm et al. (2007). Values of all diversity indices were comparable between two rivers and seasons observing significant variations both in space and time. The incongruities may be occurred due to nutrients dissimilarity (Huh and Kitting, 1985), water currents and environmental incidents (Keskin and Unsal, 1998) and fish migrations (Ryer and Orth, 1987) as well as seasonal difference in species diversity. In Bangladesh, a number of fish species breed and recruit as new stocks in their habitat from April to May that would be another rationale to alter diversity indices (Hossain et al., 2012).

Besides, present study found almost same similarity in case of occurrence of finfish assemblage between space and spell where main donating species were also similar but their percentage of contribution varied from each other. At this point, resemblance was found in more among rivers rather than seasons where major causal fishes were related to the Chalan beel for $P$. sophore and P. ticto (Kostori et al., 2011) while was disparate with Meghna River estuary (Hossain et al., 2012). The non-metric multidimensional scaling (nMDS) composed of associations among assemblages in particular coordination rooted in their similarity or dissimilarity. The spatiotemporal fish assemblage and structure of these rivers (stress as 0.24 ) just above the minimum values of nMDS $(<0.15)$ model indicate improved fitting with petite suspects by lessening the stress supported by Kruskal and Wish (1984) and Sanches et al. (2016) in relative abundance for Brazilian reservoir (stress as 0.20). Base on ANOSIM and PERMANOVA analysis, Tangon River showed dissimilarity in fish assemblage with Kulik River during monsoon may be due to have more different ecosystems. Fish assemblage was dissected among habitats and seasons would be attributable to
particular ecological variables for breeding, feeding, rearing and sheltering fluctuated seasonally with water quality parameters (Agostinho et al., 2008). In addition, the Tangon River was isolated from the Kulik River in rainy season may be resulted as more suitable breeding and feeding habitat or due to seasonal variations of ecological factors. Seasonality also controls the spawning activity of fishes accelerating to alter catch composition (van Overzee and Rijnsdorp, 2015).

In conclusion, spatiotemporal connection between fish species and ecological factors play a causative role to complete the life cycle of fishes ( $A$. jaya, $P$. ticto, $P$. sophore, $C$. gongota and $C$. punctatus). Water depth and water temperature are main leading factors altering fish assemblage and structure spatiotemporally. So, the government of Bangladesh, scientists and other authorities should recompense their attentions to protect the indigenous fishes essentially threatened species during dry or winter season in these rivers. Through further research, monitoring and raising awareness among local people adjacent these rivers, rare fishes would be regenerated and dominant species may be sustained at the Tangon and Kulik Rivers in Thankurgaon district of Bangladesh.

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